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The effects of downscaling method on the variability of simulated watershed response to climate change in five U.S. basins

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Abstract

Simulations of future climate change impacts on water resources are subject to multiple and cascading uncertainties associated with different modeling and methodological choices. A key facet of this uncertainty is the coarse spatial resolution of GCM output compared to the finerresolution information needed by water managers. To address this issue, it is now common practice to apply spatial downscaling techniques, using either higher-resolution regional climate models or statistical approaches applied to GCM output to develop finer-resolution information for use in water resources impacts assessments. Downscaling, however, can also introduce its own uncertainties into water resources impacts assessments. This study uses watershed simulations in five U.S. basins to quantify the sources of variability in streamflow, nitrogen, phosphorus, and sediment loads associated with the underlying GCM compared to the choice of downscaling method (both statistically and dynamically downscaled GCM output). We also assess the specific, incremental effects of downscaling by comparing watershed simulations based on downscaled and non-downscaled GCM model output. Results show that the underlying GCM and the downscaling method each contribute to the variability of simulated watershed responses. The relative contribution of GCM and downscaling method to the variability of simulated responses varies by watershed and season of the year. Results illustrate the potential implications of one key methodological choice in conducting climate change impacts assessments for water - the selection of downscaled climate change information.

Keywords

climate change; downscaling; streamflow; water quality; variability

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1.0 Introduction

Scenario analysis using general circulation model (GCM) output to drive hydrologic models is a common approach for assessing the potential effects of climate change on water resources. These studies are complicated by two challenges: (1) the large uncertainties associated with GCM simulations of future climate change, particularly for precipitation (e.g., see Cox and Stephenson 2007, Raisanen 2006, Stainforth *et al.* 2007, Hawkins and Sutton 2011) and (2) the coarse spatial resolution of GCM output that does not incorporate local topographic effects compared to the finer-resolution information needed by water managers (e.g. Fowler *et al.* 2007). Progress has been made with the first challenge by adopting approaches such as use of an ensemble of model runs to capture the range of variability across multiple GCMs (e.g. Tebaldi and Knutti 2007). Exploring the full range of variability in this way can reveal system vulnerabilities and guide risk management (Wilby *et al.* 2012, Weaver *et al.* 2013).

To address the second challenge, it is common practice to apply spatial downscaling methods (DSMs), using either higher-resolution regional climate models (dynamical DSMs) or statistical approaches applied to GCM output. Dynamical DSMs use GCMs to drive nested regional-scale, numerical models at higher spatial resolution to simulate local conditions in greater detail (Elguindi and Grundstein 2013, Pryor et al. 2012, Mearns et al. 2009, 2013). Statistical DSMs are based on relationships that interpolate large-scale GCM output to observations of historical weather and climate (Abatzoglou and Brown 2012, Burger et al. 2012, Wood et al. 2004, Maurer et al. 2009). Downscaling yields information at a finer spatial resolution more appropriate for watershed analysis. However, the process can also modify and/or compound the uncertainties associated with the choice of a particular underlying GCM. While downscaling can improve local-scale representation of topographic effects, this may have little meaning if the GCM misplaces key features, such as the location of the jet stream or storm tracks relative to the site of interest (Hall 2014). The choice of which projected future climates, and thus specific spatial and temporal details, are used in an assessment has a direct influence on results. It is important for practitioners to understand the potential implications of this methodological choice – the choice of a DSM – on assessment results.

Previous studies have assessed the sources of variability in simulations of hydrologic response to climate change. Chen *et al.* (2011a, 2011b) evaluated the sources of uncertainty in hydrologic projections for the Manicouagan 5 watershed in Quebec through 2100, examining the role of emission scenarios, GCM, statistical DSM, hydrologic model structure, and hydrologic model parameter sets. They found that choice of GCM is consistently a major contributor to variability across the outputs of different simulations; however, they found that choice of DSM, as well as the GCM initial conditions, could have a comparable or even larger contribution for some hydrologic endpoints. Extension of this analysis to include downscaling using four regional climate models (RCMs) found that results from statistical downscaling and RCMs had similar envelopes of uncertainty, although the RCM methods had a larger impact for some endpoints (Chen *et al.* 2013).

Conversely, variability across simulations driven by different GCMs was more pronounced in runoff projections for major French drainage basins than among DSMs, including both statistical and dynamical downscaling with quantile bias correction (Boé *et al.* 2009). Similarly, Habets *et al.* (2013) found that GCM-related variability was the largest driver of the magnitude of climate impacts on hydrogeology in northern France. Mpelasoka and Chiew (2009) found greater variability in precipitation projections among multiple GCMs than among three statistical DSMs. Finally, in hydrologic simulations of the Alpine Rhine using statistical DSMs, the choice of GCM was the dominant contributor to inter-simulation variability in summer and fall, but the choice of DSM was of greater importance in winter and spring (Bosshard *et al.* 2013). These studies suggest the relative contributions of GCM vs. DSM to variability differs among locations, but direct comparison across sites is complicated by differences in methods. The studies discussed above focus on water quantity (e.g., streamflow) and do not consider water quality endpoints.

In this study, we use watershed model simulations in five U.S. watersheds to illustrate the effects of DSM on the variability of simulated streamflow and water quality (nitrogen, phosphorus, and suspended solids) responses to climate change. Watershed model simulations are driven by meteorological inputs representing mid-21st century climate developed from non-downscaled and downscaled GCM output. Our analysis addresses two questions: (1) What is the relative contribution of GCM, DSM, and interannual variation on simulated watershed responses? (2) How do simulated watershed responses change when driven by downscaled versus non-downscaled output from a single underlying, or "parent" GCM (hereafter referred to as the incremental effects of downscaling)?

The first question allows us to explore overall sources of variation within the ensemble of simulated future changes in climate evaluated in these watersheds, while the second allows us to address three sub-questions: (a) Does the variability (i.e., range) of simulated watershed responses to climate change differ when driven by downscaled versus non-downscaled GCM information? (b) Does using downscaled data lead to the identification of regional patterns of streamflow and water quality variability not found using non-downscaled GCM output, e.g., small scale orographic effects? (c) Does the simulated watershed responses to climate change depend on the particular GCM and/or type of downscaling used?

Results illustrate the potential implications of one key methodological choice in conducting climate change impacts assessments for water – the selection of DSM. Other known sources of variability including watershed model structure and parameters (e.g., see Mendoza *et al.* 2014), emissions scenarios, or other factors in the "uncertainty cascade" (e.g., see Wilby and Dessai 2010) are not evaluated. Our intent is to help bridge gaps between the climate and hydrologic modeling communities and improve the integration of modeling efforts across these communities (Lofgren and Gronewald 2013).

2.0 Methods

Our analysis is based on simulations of five large watersheds: the Minnesota River watershed, the Apalachicola-Chattahoochee-Flint River (ACF) watersheds, the Willamette

River watershed, the Salt River watershed, and the Susquehanna River watershed (Figure 1). All watershed simulations used in the analysis were conducted as part of a larger, previous modeling effort to assess streamflow and water quality sensitivity to climate change in 20 U.S. watersheds (USEPA, 2013; Johnson *et al.* 2012). Except for the Salt River watershed, the study watersheds are comparable in size to the United State Geological Survey's (USGS) Hydrologic Unit Code (HUC) 4-digit basins, ranging from 15,025 km² (Salt) to 71,236 km² (Susquehanna), and were selected to represent different hydroclimatic and watershed conditions occurring throughout the nation (Table 1).

Watershed simulations were conducted using the Soil and Water Assessment Tool (SWAT – version 2005; Neitsch *et al.* 2005). The SWAT watershed model incorporates data for weather, soils, topography, vegetation, and land use and cover to estimate water and sediment movement, nutrient cycling, and other watershed processes in large, complex watersheds (Neitsch *et al.* 2005). Potential evapotranspiration (PET) was calculated internally in SWAT using the Penman-Monteith energy balance method (Allen et al. 2005). Land use and land cover was from the 2001 National Land Cover Database (NLCD) and held constant in all simulations to focus on the effects of climate change and DSM. In each of the five study watersheds, SWAT was used to simulate changes in total streamflow, total nitrogen (TN), total phosphorus (TP), and total suspended solids (TSS) loads in response to simulated mid-21st century climate change.

SWAT models for each study watershed were calibrated and validated at the scale of USGS 8-digit HUCs. All models performed credibly for hydrology with total volume errors within 20 percent and Nash-Sutcliffe coefficients of model fit efficiency for monthly streamflow ranging from 0.32 to 0.83. Confidence limits (95%) on mean monthly flows at downstream gages ranged from $\pm 3\%$ (Susquehanna) to $\pm 15\%$ (Salt). Water quality simulation focused on monthly loads and has much higher uncertainty due to limited availability of sampling data. In most cases, however, the pollutant load simulations from SWAT models generally appear to be in the fair to good range (median absolute error of 16.5% relative to loads estimated from sparse monitoring data). All analyses in this study are based on simulation results expressed as mid-21st century changes relative to historical baseline conditions. The setup, calibration, and validation of SWAT models in each of the five study watersheds is described in detail in the appendices to U.S. EPA (2013). Simulations use consistent methods, models and scenarios to facilitate comparison among study watersheds.

2.1 Simulated future climate change

All projected future climates are based on mid-21st century (2041–2070) climate model simulations using the four GCMs from Phase 3 of the Coupled Model Intercomparison Project (CMIP3) under the A2 emissions scenario (IPCC, 2007) covered by the regional downscaling efforts of the North American Regional Climate Change Assessment Program (NARCCAP; http://www.narccap.ucar.edu). The specific future climate information used to drive watershed simulations differs depending on if or how GCM output was downscaled. We consider three categories of climate change information based on these same underlying GCMs: non-downscaled GCMs, dynamically downscaled NARCCAP projects, and statistically downscaled Bias Corrected and Spatially Disaggregated (BCSD) (Maurer *et al.*

2009) projections (Table 2). These three categories of climate change information, while not comprehensive of all GCMs or DSMs, are representative of commonly applied 'off the shelf' datasets used in climate change impacts studies.

The NARCCAP information is dynamically downscaled using RCMs nested within parent GCM models to represent detailed sub-grid, regional processes, and is intended to provide greater detail at finer spatial resolution than the driving GCM. NARCCAP RCM output is spatially downscaled to a 50×50 km grid over North America (Mearns *et al.* 2009; Mearns *et al.* 2007). This downscaled output is archived for two 30-year time slices (1971–2000 and 2041–2070) at a temporal frequency of three hours. All the NARCCAP simulations assume the IPCC's A2 greenhouse gas storyline (IPCC 2007). We evaluated six NARCCAP GCM/RCM combinations (U.S. EPA, 2013).

The BCSD information is statistically downscaled as described by Wood et al. (2004) and Maurer *et al.* (2007). This dataset provides temperature and precipitation on a 1/8-degree (approximately 14×10 km at 45 °N) horizontal grid. We evaluate four BCSD-derived future climates based on the same four underlying GCMs used by NARCCAP. For consistency with the NARCCAP scenarios, we use the 2050 CMIP3 BCSD scenarios for the A2 emissions storyline.

Finally, non-downscaled future climate projections are based directly on GCM output. We evaluate four projected future climates based directly on the four parent GCM output used by NARCCAP and BCSD. (Note, however, that many of the CMIP3 GCMs ran multiple versions of the A2 simulation, differing only in initial conditions, to better capture the random internal variability of the climate system and to extract a more robust signal of the anthropogenic climate forcing. These multiple versions are called 'ensemble members', and the BCSD data we used derived from the HadCM3 and CCSM GCMs are from different ensemble members compared to the corresponding non-downscaled or NARCCAP data derived from the same GCM. By contrast, for CGCM3 and GFDL, the ensemble member used is identical across BCSD, NARCCAP, and non-downscaled GCM samples.)

The climate change information (e.g., from NARCCAP, BCSD, and non-downscaled GCM) used to drive SWAT watershed models in each study watershed was implemented as daily meteorological time series. In each case, daily time series were created using the "change factor" or "delta" method (e.g., see Anandhi *et al.* 2011). The change factor method combines information about relative change (between a historical period and future period, generally of a number of years or decades in length) in a particular climate variable of interest, such as temperature or precipitation, with one or more observed local time series of the same variable, to create a synthetic future input dataset for (in this case) the SWAT model.

Climate model outputs were bi-linearly interpolated to each of the NCDC weather stations used by the SWAT models (see Table 1 for the number of stations in each watershed). Monthly change statistics (change factors) for each of the 14 total sources of future climate information (from NARCCAP, BCSD, or the non-downscaled GCM output) at each weather station were then calculated as the difference between mid-21st century (2041–2070) and

It is important to note that the different DSMs do not all provide the same meteorological variables. SWAT watershed simulations in this study estimate potential evapotranspiration (PET) using the Penman-Monteith energy balance method, which requires inputs for solar radiation, humidity, and wind. To provide a consistent basis for comparison, simulated future climate change in this study represents changes only in air temperature and precipitation; the only two variables commonly archived for each DSM. Other climate variables needed to compute PET by the energy balance method are left unperturbed in this study as supplied by SWAT's weather generator representation of existing climate. Accordingly, information about potential future change represents the effects of changes in air temperature on PET, but does not account for changes in solar radiation, humidity, and wind. This de-linking of mass inputs (precipitation) and energy inputs other than average air temperature is a simplification, but reflects common practice in many climate impact studies (Milly and Dunne 2011). Note that results for SWAT simulations of these watersheds reported in U.S. EPA (2013) and Johnson et al. (2015) do make use of projected changes in these energy inputs where available.

2.2 Data Aggregation and Analysis

used.

SWAT simulations in each study watershed resulted in 29-30 years of daily output for each future climate simulation evaluated. Daily output was first aggregated to time series of annual and seasonal averages. Aggregated values within each study watershed were then normalized by the mean and standard deviation of their baseline scenario (1971-2000). This converts each time series to a set of deviations from mean baseline conditions that share a common scale of projected change across watersheds. Our analysis focuses on projected changes in endpoint values. Changes were calculated by subtracting baseline deviations from future climate deviations. The endpoints we consider are total streamflow, total nitrogen load (TN), total phosphorus load (TP), and total suspended solids load (TSS). All analyses were evaluated, using mixed effects models with restricted maximum likelihood, at seasonal and annual intervals with the 'lme4' package in R (R Core Team, 2014; Bates et al. 2014).

Mixed effects models (i.e. hierarchical or multilevel models) can be used when data are nested within groups or categories, such as the climate models in this study (Zuur et al. 2009). These models contain both fixed and random effects, where the fixed effects evaluate overall, population-level relationships, and the random effects account for and produce estimates of heterogeneity among the groups or categories for the fixed effects. Like classical ANOVAs that can incorporate random effects, e.g., those used to analyze randomized block designs, the goal is not to evaluate differences between the groups but variability.

2.2.1 Analysis of Sources of Hydrologic Variation within the Ensemble— Mixed effect models were used to quantify the overall variability associated with parent GCM and DSM in the ensemble of 14 sets of simulated future climate for streamflow and water quality endpoints in each of the five study watersheds. In these models, we used the parent GCM (GCM, four groups; columns of Table 2) and DSM (three or four groups per parent GCM; cells of Table 2) as categorical random factors, with the DSM factor nested within the GCM factor. The mixed effects models produced estimates of the mean projected ensemble change (β_0) and of three sources of variation: variation among parent GCMs (σ_{GCM} ,), variation among DSMs within parent GCMs ($\sigma_{DSM/GCM}$), and the unaccounted for interannual variability within DSMs ($\sigma_{residual}$). The approach is analogous to a traditional nested analysis of variance (ANOVA), except it produces estimates of variation for the two random factors instead of mean GCM and DSM estimates. These effects are illustrated in Figure 2. The three estimated standard deviations produced by these statistical models are useful because they allow us to visualize the distribution of projected changes across the full ensemble of simulations with respect to both the parent GCMs and DSMs.

2.2.2 Analysis of Incremental Effects of Downscaling GCM Output on

Hydrologic Simulations—Analyses were also conducted to assess how simulated watershed responses change within the overall ensemble of simulations, when driven by downscaled versus non-downscaled GCM output (hereafter referred to as the incremental effects of downscaling). First, we used a mixed effect model with two fixed, binary, categorical variables to compare non-downscaled GCMs (controls) to downscaled means: BCSD, yes (1) or no (0); and NARCCAP, yes (1) or no (0). To measure the variability in these effects, we allowed the relationship between non-downscaled GCMs and downscaled means to vary randomly across parent GCMs (four groups; columns of Table 2). Because the "NARCCAP" variable in this statistical model does not distinguish between the two RCMs associated with CGCM3 and GFDL (e.g., both CGCM3-CRCM and CGCM3-RCM3 would have the same covariate values), we evaluated each mixed effects models four times, once with each unique combination of CGCM3 and GFDL GCMs and RCMs (Table 2) and then reported average parameter estimates and p-values derived from those averages.

Each statistical model produced estimates of the mean projected change among the group of non-downscaled GCM data (β_0) and the difference between that and the mean projected change in the group of BCSD and NARCCAP data (β_{BC} and β_{NR}). These differences are illustrated in Figure 3a. These models also produce estimates of four sources of variation (standard deviations): variation among non-downscaled GCMs(σ_{GCM}), variation in the effect of downscaling with BCSD among parent GCMs(σ_{BC}), variation in the effect of downscaling with different NARCCAP projections among parent GCMs(σ_{NR}), and the unaccounted for interannual variation ($\sigma_{residual}$). This approach is analogous to a traditional analysis of covariance (ANCOVA) with two covariates and four groups (the parent GCMs), except it produces estimates of variation among the groups for the fixed parameters (β_0 , β_{BC} , and β_{NR}), instead of four intercepts associated with individual non-downscaled GCMs and eight projected changes associated with individual BCSD and NARCCAP projections.

To visualize these random effects consider Figure 3b and 3c. Figure 3b shows differences between individual NARCCAP projections and their associated non-downscaled GCM projections (dashed red lines), compared to the overall group difference (thick black line). Figure 3c shows how those differences are represented in the model. Orange arrows show how each non-downscaled GCM projection compares to the overall group. The mixed effects model estimates the variability associated with those differences (σ_{GCM}), while a traditional ANCOVA would estimate the magnitude of each difference separately. Blue arrows show how the difference between each NARCCAP projection and their associated non-downscaled GCM projection compares to the overall group difference (thin black lines have been added to highlight this comparison).

The mixed effects model estimates the variability in these downscaling effects (σ_{BC} and σ_{NR}), while a traditional ANCOVA would estimate the magnitude of each effect separately. Of these sources of variation, we are interested in the variability in the BCSD and NARCCAP effect (σ_{BC} and σ_{NR}). When significant, BCSD or NARCCAP fixed effects (β_{BC} and β_{NR}) indicate that the application of downscaling consistently found regional patterns not found in the non-downscaled GCM output, resulting in directional shifts in simulated streamflow or water quality responses. Significant BCSD or NARCCAP random effects (σ_{BC} or σ_{BC}) indicate that the magnitude or direction of the BCSD or NARCCAP random effects (σ_{RC} or σ_{BC}) indicate that the magnitude or direction of the BCSD or NARCCAP and model effects depends on the parent GCM and downscaling model: the larger the value of σ_{BC} or σ_{NR} , the larger the discrepancy between the overall BCSD or NARCCAP effect and individual model combinations.

We then used three simple mixed effects models to estimate variability of watershed simulations using BCSD and NARCCAP downscaled climate to compare against the variability among non-downscaled GCMs. This differs from the previous analysis in that it allows us to visualize the variability among the three categories of climate simulations separately. In these statistical models, either the non-downscaled GCM, BCSD, or NARCCAP (four groups per category; rows of Table 2) was used as a categorical random effect. Here also we evaluated each mixed effects models four times, once with each unique combination of CGCM3 and GFDL GCMs and RCMs (Table 2).

These statistical models produced estimates of the mean projected change for each group of projections, the non-downscaled GCM, BCSD, or NARCCAP projections and of two sources of variation: variation among DSMs (σ_{GCM} , σ_{BCSD} , or $\sigma_{NARCCAP}$) and the unaccounted for inter-annual variation within the DSMs, ($\sigma_{residual}$). This approach is analogous to a traditional one-way ANOVA, except that it produces estimates of variation instead of mean non-downscaled GCM, BCSD, or NARCCAP responses. Figure 3d can be used to compare these random effects to the previous model. Orange arrows again show how each non-downscaled GCM projection compares to the overall group of non-downscaled GCMs. Both models produce a valid estimate of σ_{GCM} , but for consistent comparisons we report the version estimated here. Green arrows, however, show how each NARCCAP projection compares to the overall group of variation, we are interested in the variability among DSMs for each group (σ_{GCM} , σ_{BCSD} , or $\sigma_{NARCCAP}$). Taking all parts of Figure 3 together, the group of NARCCAP scenarios estimates a larger

hydrologic response to climate change than the non-downscaled GCMs, but the variability in the NARCCAP effect also leads to larger variability among the NARCCAP scenarios.

3.0 Results and Discussion

Simulated changes in streamflow and water quality endpoints in response to the 14 different projected future climates in each of the five study watersheds are shown in Figure 4. The values shown are ratios of future (2041–2070) to baseline (1971–2000) annual average values at the downstream outlet of each study watershed. Symbols represent watershed responses to climate change based on non-downscaled GCM, NARCCAP, and BCSD data. Projected average changes in air temperature, precipitation, actual evapotranspiration (AET), and potential evapotranspiration (PET) for each climate future are also shown for comparison. Figure 4 illustrates a wide range in simulated water quality endpoints when different categories of future climate change information are used to drive SWAT. The range of water quality responses is generally wider (on a percentage basis) than the range in driving climate variables, and in most cases spans unity (indicating disagreement about the sign of future change). This reflects the cascading effects of variability in climate drivers when coupled with watershed modeling to assess watershed responses. In addition, Figure 4 shows the important role of water limitation in certain regions and seasons of the year, as revealed by the difference in future change between actual and potential evapotranspiration.

3.1 Sources of Variation within the Ensemble

Analysis of simulation results in the five study watersheds show that parent GCM, and DSMs within each parent GCM, can each be a significant source of variability in the overall ensemble of projected streamflow and water quality responses to climate change. The relative contribution of GCM and DSM to the variability of simulated streamflow and water quality endpoints, however, varies by watershed, season of the year, and streamflow and water quality that cannot be attributed to GCM and DSM also varies among watersheds, season, and endpoint. Parameter estimates for all models are presented in Table S1.

Results show differences in the hydrologic and water quality response to climate change among the five study watersheds. This is expected due to differences in watershed physiographic, hydroclimatic, land use and other factors. Hydroclimatic conditions vary from the arid southwest (Salt) to the humid Pacific Northwest (Willamette) and Southeast (ACF) and represent both continental and maritime mid-latitude climates. For example, in ACF and Minnesota the estimated variability among parent GCMs was most often smaller than the variability of DSMs within the parent GCMs (Figure 5). In ACF the variability among parent GCMs was smaller than the variability among DSMs within parent GCMs in all models considered, while in Minnesota this was true in 80% of the models considered. In contrast, variability among parent GCMs in 75% of the models, suggesting that either large or small differences exist between parent GCMs or DSMs, respectively. Much of the variability we observe across regions may depend on simulated precipitation and spring warming, as the timing and spatial distribution of precipitation has been shown to vary widely across

climate models, which in topographically complex watersheds, or those that are influenced by small-scale meteorology, can result in very different flow patterns (Rasmussen *et al.*, 2012).

Differences in the relative contribution of GCM and DSM among study watersheds can be illustrated by comparing results for the ACF and Willamette basins (Figure 6), which have different hydroclimatic and watershed attributes. For Willamette, GCMs tend to be more important than DSMs in determining variability for streamflow and water quality endpoints, while the reverse is true for ACF. Contributing to this difference, Willamette is strongly influenced by the large-scale flow (e.g., the North Pacific storm track) year round, particularly in the cold season, over which the choice of GCM would be expected to play a larger role. By contrast, temperature and precipitation in ACF strongly depend on smaller-scale meteorology (e.g., local convection) that DSMs (particularly dynamical downscaling) would be more likely to resolve. In addition, the Willamette is closer to the inflow boundary of the RCM domains, so it is likely more strongly influenced by the driving GCM solution, whereas regional climate simulated at ACF experiences more modification as the meteorological flow traverses the RCM domains.

These results are consistent with Wang *et al.* (2009), who compared the performance of six RCMs over the Intermountain Region of the Western U.S. to data from the North American Regional Reanalysis (NARR) dataset and demonstrated that the different RCMs are largely consistent in the Cascade Range (OR, WA) where the dominant upper level flow first encounters land. The differences among RCMs reported by Wang *et al.*, and the difference from NARR, are greatest on the windward side of the Rocky Mountains in Colorado and remain large into Arizona (location of the Salt watershed).

Simulations within each of the five study watersheds also show differences in the relative contributions of GCM and DSM in different seasons of the year. In our ensemble of projected future climates, variability among parent GCMs was smaller than the variation among DSMs within parent GCMs most often in autumn and winter (Figure 5). The variability among parent GCMs was always smaller than among DSMs in winter, while in autumn it was smaller in 65% of the models considered. The converse was true in spring, where variability among parent GCMs was larger than among DSMs in 65% of the models. Projected changes that used downscaled results tended to deviate from non-downscaled results most in winter (discussed below, Figures 7-11). Bosshard et al. (2013) also found that DSM contribution to variance was larger during winter months. Apart from these patterns in the larger data set, each watershed had their own unique characteristics driven by its hydroclimatic setting (Figure 5). For example, variability in projected changes tended to be highest in spring for Willamette, summer for Salt, autumn for ACF and Susquehanna, and winter for Minnesota. Specifically, Willamette has relatively high mountains where spring snowmelt is important. The Salt is affected by summer monsoons, and ACF has highly variable tropical storms in late summer and fall. Winter has the highest variability in Minnesota, likely in part because scenarios resolve winter temperatures and the difference between precipitation as rain or snow differently.

Simulations show less pronounced differences in the relative contributions of GCM and DSM for different streamflow and water quality endpoints. The variability of streamflow and water quality endpoints is most pronounced for the Minnesota and Salt Rivers (Figures 4, 5). The relationship between GCM and DSM effects across endpoints was relatively consistent, but inter-annual variation that could not be attributed to each source varied widely by metric. Unaccounted for interannual variation in streamflow was larger than the other two effects in only 32% of the models, but this value increases to 68%, 80%, and 96% in TSS, TN and TP models. These results illustrate the greater variability in projected changes in water quality metrics, especially TN and TP, due to multiple interacting factors affecting pollutant sources, fate and transport, such as changes in precipitation intensity and seasonal timing relative to plant growth cycles.

3.2 Incremental Effects of Downscaling GCM Output on Hydrologic Simulations

The incremental effects of downscaling were evaluated by comparing SWAT simulations in the five study watersheds when driven by downscaled versus non-downscaled climate change information from the same parent GCM. By "incremental effects", we mean the quantified impacts, on simulated hydrologic endpoints, of using dynamical or statistical downscaling to modify the output from a given GCM. This is distinct from the overall variability among GCMs and/or DSMs, as presented in the previous section.

Figures 7 -11 show results for each streamflow and water quality endpoint by season of the year. The significance of fixed and random BCSD and NARCCAP effects are also shown in Figures 7–11. Parameter estimates for all of the effects models are presented in Table S2 and S3. Results show significant variability in the effects of downscaling among watersheds, seasons of the year, and to a lesser extent with the different streamflow and water quality endpoints. In some cases (e.g., watersheds/seasons), watershed simulations driven by downscaled (BCSD, NARCCAP) versus non-downscaled (GCM) climate change information deviate in a consistent direction, suggesting that downscaling is capturing some common underlying process in the watershed, e.g., orographic effects or lake snow, that the GCMs are not. In other cases, however, simulations using NARCCAP versus BCSD deviate from the GCM in ways that are not consistent with each other, including the sign of the projected change (e.g., recall Figure 4 and see discussion in Johnson et al., 2012).

Simulations within individual study watersheds tend to show greater incremental effects of downscaling when driven by climate change information from NARCCAP RCMs (i.e. significance and size of the fixed BCSD and NARCCAP effects, 16% vs. 26% overall, Figures 7–11). In many cases the variability among BCSD and NARCCAP scenarios was similar, but random NARCCAP effects were significant more often (64% vs. 96% overall). In other words, for the ensemble of projected future climates in this study, NARCCAP RCMs were on the whole more likely to find consistent regional patterns that differed from non-downscaled GCMs, but individually these differences were more variable. This result could occur because, unlike with statistical downscaling, RCMs are able to alter the atmospheric circulation and convective environment in the parent GCM.

Simulations across the five study watersheds show significant variability in the incremental effects of downscaling in these different hydroclimatic and physiographic locations. For

example, looking across all streamflow and water quality endpoints and seasons of the year, watershed simulations driven with downscaled climate change information (i.e., NARCCAP and BCSD) differed from simulations using non-downscaled GCMs most often in the Salt watershed. In the Salt, simulations using BCSD differed significantly from the non-downscaled GCM runs in 20% of the simulations, while those based on NARCCAP differed significantly in 60% of the simulations (Figures 7–11). With the exception of the summer season, which had highly variable changes in streamflow (Figures 5, 10), the use of climate change information from BCSD in the Salt resulted in relatively higher streamflow and loads, while the use of NARCCAP data resulted in relatively lower streamflow and loads. This is contrasts with the ACF and Willamette watersheds, where in the former, simulations driven with BCSD differed significantly from the non-downscaled GCMs in 30% of the cases (and simulations with NARCCAP did not differ), while in the latter, simulations driven by NARCCAP differed significantly in 20% of the cases (and simulations with BCSD did not differ). Across all five study watersheds, however, random NARCCAP effects were significant more often than random BCSD effects (Figures 7–11).

Lastly, Figures 7–11 illustrate variability in the incremental effects of downscaling when considering different streamflow and water quality endpoints, and seasonal differences in endpoint values throughout the year. While the effects of downscaling were relatively consistent among the different endpoints, results are more variable across seasonal endpoint values. For example, considering annual average streamflow, simulations driven by downscaled climate change information from BCSD and NARCCAP differed significantly from those using non-downscaled GCMs in 5% and 35% of models, respectively (Figure 7). A similar pattern of significant fixed effects occurs in the autumn and summer seasons, where BCSD and NARCCAP effects differed from non-downscaled GCMs in 5% and 25% of models in autumn, and 0% and 35% of models in summer, respectively (Figures 10 and 11). In the spring and winter seasons, however, BCSD and NARCCAP effects differed in 20% and 0% of models in spring, and in 50% and 30% of models in winter, respectively. In other words, BCSD effects were most significant in spring and especially winter (20% and 50% of models, respectively), while NARCCAP effects were significant in roughly equal proportions (25–35% of models) for all periods except spring.

3.3 Assumptions and Research Needs

This study describes a particular set of watershed simulations to illustrate how driving a watershed model with different approaches to downscaling climate change information can influence simulation results. All results are conditional on the methods, models, and climate change information evaluated in the underlying simulations. Several caveats should be noted. First, to provide a consistent basis for comparison, all simulations of watershed response to climate change assume future changes only in air temperature and precipitation. We intentionally do not consider the implications of representing changes in other meteorological variables such as humidity, radiation and wind speed that are necessary to calculate PET using an energy balance approach (e.g., see Milly and Dunne 2011). Representation of these additional meteorological variables can have a significant influence on watershed simulation results. Sensitivity studies in the five study watersheds suggest inclusion of projected changes in dewpoint resulted in a reduction in estimated annual PET

of about 11% across all the meteorological stations, implying an underestimation of soil moisture and streamflow when change in dewpoint is not used (see USEPA 2013). The importance of accounting for dewpoint for properly simulating hydrology under future climate change was also noted by Pierce et al. (2013).

In addition, all meteorological inputs used to drive the watershed models were created using the change-factor method applied to historical time series. Use of change factors to translate climate change information to site-specific information to drive watershed models is, in itself, a simple additional step of statistical downscaling from the gridded DSM output to point gauge locations. Chen *et al.* (2013) have shown that different approaches for translating regional climate projections to site-specific inputs for hydrologic models can impact watershed simulations. In this study, we do not consider the implications of using different types of change factors (e.g., scaling versus quantile mapping), nor do we compare the change factor application to approaches that use the climate model-simulated sequences of precipitation events.

Finally, it must be noted that simulations of potential climate change impacts are subject to multiple and cascading uncertainties associated with different watershed model characteristics and methodological choices. In this study we address only one source – the effects of downscaling climate change information used to drive watershed models – and do not address other uncertainties affecting watershed simulations. The analysis is nevertheless illuminating and shows promise for providing systematic, quantitative uncertainty characterization in the study of watershed responses to climate change.

Future research addressing the above and related methodological questions would be valuable. In addition, application of this type of statistical analysis to additional study areas, increasing the sampling across diverse hydroclimatic regimes, would be helpful for eliciting clearer patterns in the relative importance of GCM and DSM by watershed characteristics. Increasing the size of the GCM ensemble, and therefore range of future climates, considered might similarly produce insights into more systematic patterns of response in watershed simulations to GCM versus DSM forcing. Finally, while the kind of "ensemble of opportunity" approach to pairing of GCMs and DSMs we have used here allows for the leveraging of a large volume of existing projection data, it makes it difficult to separate variability in hydrologic endpoints due simply to the increased resolution from that due to factors such as RCM model formulation or the particular statistical downscaling algorithm used. It would therefore be worthwhile to repeat this type of analysis in the context of "big brother" or "perfect model" approaches (e.g., see Denis et al. 2002), where a high-resolution climate simulation is degraded to coarser resolution to create a synthetic analogue of both the GCM and "downscaled" data from the same underlying model run.

4.0 Summary and conclusions

Assessments of climate change impacts on water resources are complicated by the scale, complexity and inherent uncertainty of the problem. This study illustrates one poorly understood but important facet of this complexity; the potential effects of DSM (including the choice to use downscaling at all) on simulations of hydrologic and water quality

changes. Our results show that both the parent GCM and how downscaling is done can contribute to the variability of projected watershed responses. Moreover, sources of variability differ among watersheds, season of the year, and for different streamflow and water quality endpoints governed by different watershed and hydroclimatic processes. The differences among GCMs can be the major source of variability in some cases, while if and how the data are downscaled can be a major factor in others. Our results also provide a detailed illustration of how downscaling GCM output can alter simulations of watershed processes as compared to simulations based on non-downscaled GCMs. Water resources practitioners should be aware that while models are a useful and necessary part of management planning, there is significant uncertainty in projections associated with both GCM choice and DSM choice. Given the uncertainties, managers should seek to examine a wide range of plausible futures, identify potential vulnerabilities, and focus on solutions that are robust across a range of plausible futures rather than a single most likely future.

Statistical downscaling has power in its ability to reproduce local-scale deviations from areal average results, such as finer-scale orographic effects, and can adjust for some inherent spatial biases in GCMs, but it assumes historical spatial relationships between GCM output and local climate will remain unchanged over time. Statistical downscaling is also less computationally intensive and thus more conducive to running larger ensembles of scenarios. Dynamical downscaling with RCMs is a physics-based approach that attempts to account for changes in the relationship between global and local climate, but requires a high level of effort and is not yet proven to yield more credible results. There is no consensus on a "best" downscaling approach for use in assessment of climate change impacts on water resources. Statistical and dynamical methods each have advantages and disadvantages, and there are a wide variety of specific methods within each category. In choosing information sources for potential future climate change, one should consider the study goals and specific questions being asked, level of confidence required for information to be actionable, time and resources available, and other relevant questions that determine the decision context.

Supplementary Material

Refer to Web version on PubMed Central for supplementary material.

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Nover et al.



Figure 1.

Location of the five study watersheds.

mean projected ensemble change (β_0)



Figure 2.

Illustration of model terms (sources of uncertainty) used in our analysis. The model estimates one fixed effect, the mean ensemble projected change (β_0). The variation among parent GCMs (orange arrows) is estimated with σ_{GCM} , the variation among DSMs within parent GCMs (blue arrows) is estimated with $\sigma_{DSM/GCM}$; and the residual, or unaccounted for inter-annual variation within DSMs (green arrows) is estimated with $\sigma_{residual}$.

Nover et al.



Figure 3.

Illustration of the "incremental effects" of downscaling GCM data in our analysis. (a) The model estimates three fixed effects, the mean projected change in the group of non-downscaled GCM scenarios (β_0), and the difference between that and the mean projected change in the group of BCSD and NARCCAP scenarios (thick black arrows, β_{BC} and β_{NR}). (b) As an example, three NARCCAP scenarios (red dashed arrows) are used to illustrate two sources of variability in this model. (c) Here the variation among non-downscaled GCMs (orange arrows) is estimated with σ_{GCM} and the variation in the NARCCAP effect among parent GCMs (blue arrows) is estimated with σ_{NR} . The thin parallel black lines are placed to show the difference in slope between the three scenarios (green arrows) is estimated with $\sigma_{NARCCAP}$ and compared to the variability among non-downscaled GCMs (σ_{GCM}).



Figure 4:

Simulated future changes in climate, streamflow, and water quality endpoints in the five study watersheds. Except for temperature, points are the ratio of future (mid-twenty first century) to baseline mean annual values for non-downscaled GCM (O), BCSD (), and NARCCAP (+) scenarios. Absolute difference is shown for temperature.



Figure 5:

Variability of simulated streamflow and water quality endpoints contributed by GCM, DSM, and inter-annual variability of climate change scenarios. Measures of variation among parent GCMs ($\hat{\sigma}_{GCM}$, orange), among DSMs within parent GCMs, ($\hat{\sigma}_{DSM|GCM}$, blue), and unaccounted for inter-annual variation ($\hat{\sigma}_{residual}$, green) in standardized projected changes in streamflow and water quality endpoints in each of the five study watersheds.

Nover et al.



Figure 6.

Comparison of GCM and DSM variance components for ACF (O) and Willamette (). Values show the proportion of variability among the scenarios attributed to the parent GCM: $\hat{\sigma}_{GCM}/(\hat{\sigma}_{GCM} + \hat{\sigma}_{DSM|GCM})$.



Figure 7:

The effects of downscaling on simulated annual streamflow and water quality endpoints in the five study watersheds. Boxes show projected mean annual changes in hydrology that used non-downscaled GCM, BCSD, and NARCCAP data. Arrows show the magnitude of change between non-downscaled and downscaled groups (and $\hat{\beta}_{\Delta NR}$), and dashed vertical lines show how variable that change can be across model combinations ($\pm \hat{\sigma}_{\Delta BC}$ or $\pm \hat{\sigma}_{\Delta NR}$). Solid vertical lines show variability in mean response among scenarios that used non-

downscaled GCM and downscaled data ($\pm \hat{\sigma}_{GCM}$, $\pm \hat{\sigma}_{BCSD}$, or $\pm \hat{\sigma}_{NARCCAP}$). Red boxes and dashed lines represent significant (P < 0.05) fixed and random BCSD and NARCCAP effects. All values are standardized projected changes.





Same as Figure 7, except for winter streamflow and water quality endpoints.





Same as Figure 7, except for spring streamflow and water quality endpoints.



Figure 10:

Same as Figure 7, except for summer streamflow and water quality endpoints.





Table 1.

Summary attributes of the five study watersheds.

Study Watershed	Total Area (km²)	Elevation Range (m above mean sea level)	Urban/ Residential (%)	Agriculture (%)	Forest (%)	Average Precipitation (cm/yr)	Average Temperature (°C)	# Meteon Sta (Temp Precip
ACF	49,943	0 - 1,325	9.3	21.6	47.9	138	17.46	23
Salt	15,025	585 - 3,477	0.6	0.02	61.1	56	12.92	8/
Minnesota	44,002	208 - 650	6.6	78.0	2.9	72	6.61	32
Susquehanna	71,236	0 – 957	7.4	27.0	61.1	105	9.03	27
Willamette	29,032	0-3,185	7.2	20.7	56.2	148	10.66	29

Table 2.

GCMs and downscaling methods (DSMs) used to create climate change scenarios.

	Parent GCM					
	CGCM3	HADCM3	GFDL	CCSM		
	None (non-downscaled)	None (non-downscaled)	None (non-downscaled)	None (non-downscaled)		
Democratics Method	Statistical (BCSD)	Statistical (BCSD)	Statistical (BCSD)	Statistical (BCSD)		
Downscaling Method	CRCM (NARCCAP)	HRM3 (NARCCAP)	GFDLhires (NARCCAP)	WRFG (NARCCAP)		
	RCM3 (NARCCAP)		RCM3 (NARCCAP)			

Note: Downscaling methods are non-downscaled GCM, statistical downscaling [Bias Corrected and Spatially Disaggregated archive (BCSD)], or RCM used for dynamical downscaling (NARCCAP). Model abbreviations are as follows: CGCM3, Third Generation Coupled Global Climate Model (http://www.ec.gc.ca/ccmac-cccma/default.asp?lang=En&n=4A642EDE-1); HadCM3, Hadley Centre Coupled Model, version 3 (http:// www-pcmdi.llnl.gov/ipcc/model_documentation/HadCM3.htm); GFDL, Geophysical Fluid Dynamics Laboratory GCM (http://www-pcmdi.llnl.gov/ipcc/model_documentation/GFDL-cm2.htm); CCSM, Community Climate System Model (http://www-pcmdi.llnl.gov/ipcc/model_documentation/GFDL-cm2.htm); CCSM, Community Climate System Model (http://www-pcmdi.llnl.gov/ipcc/model_documentation/CSM3.htm); GRCM, Canadian Regional Climate Model (http://www.ec.gc.ca/ccmac-cccma/default.asp? lang=En&n=4A642EDE-1); RCM3, Regional Climate Model, version 3 (http://www.ec.gc.ca/ccmac-cccma/default.asp? lang=En&n=4A642EDE-1); RCM3, Regional Climate Model, version 3 (http://www.ec.gc.ca/ccmac-cccma/default.asp? lang=En&n=4A642EDE-1); RCM3, Regional Climate Model, version 3 (http://wers.ictp.it/~pubregcm/RegCM3/); HRM3, Hadley Regional Model 3 (http://precis.metoffice.com/); WRFG, Weather Research and Forecasting Model, using the Grell convection scheme (http://www.wrf-model.org/ index.php); GFDL hi res, Geophysical Fluid Dynamics Laboratory 50-km global atmospheric time slice (http://www-pcmdi.llnl.gov/ipcc/ model_documentation/GFDLcm2.htm).