# Original article <br> Breakdown of the species-area relationship in exotic but not in native forest patches 

Tibor Magura ${ }^{a}$, András Báldi ${ }^{b, *}$, Róbert Horváth ${ }^{a}$

${ }^{\text {a }}$ Directorate of the Hortobágy National Park, Sumen u. 2, Debrecen H-4024, Hungary
${ }^{\mathrm{b}}$ Animal Ecology Research Group of the Hungarian Academy of Sciences and the Hungarian Natural History Museum,
Ludovika tér 2, Budapest H-1083, Hungary


#### Abstract

We studied the pattern of bird species richness in native and exotic forest patches in Hungary. We hypothetized that species-area relationship will depend on forest naturalness, and on the habitat specialization of bird species. Therefore, we expected strong species-area relationship in native forest patches and forest bird species, and weaker relationship in exotic forest patches containing generalist species. We censused breeding passerine bird communities three times in 13 forest patches with only native tree species, and 14 with only exotic trees in Eastern Hungary in 2003. Although most bird species (92\%) of the total of 41 species occurred in both exotic and native forests, the species-area relationship was significant for forest specialist, but not for generalist species in the native forests. No relationship between bird species and area was found for either species group in the forest with exotic tree species. The comparison of native versus exotic forest patches of similar sizes revealed that only large ( $>100 \mathrm{ha}$ ) native forests harbor higher bird species richness than exotic forests for the forest specialist bird species. There is no difference between small and medium forest patches and in richness of generalist species. Thus, the species-area relationship may diminish in archipelago of exotic habitat patches and/or for habitat generalist species; this result supports the warning that the extension of exotic habitats have been significantly contributing to the decline of natural community patterns. © 2008 Published by Elsevier Masson SAS.


## 1. Introduction

Forests are invaluable for human beings. They are important for recreation and human well being, and provide numerous ecosystem services, and are vital for the maintenance of the majority of biological diversity on Earth (Lacaze, 2000; Dirzo and Raven, 2003; Ozanne et al., 2003; Lewis, 2006). These roles need to be considered when we try to manage forests. For example, the use of wood for heating and timber in construction
or paper industry needs fast growing trees, planted in a regular pattern for easy management, without other tree or scrub species. These plantations often are of non-native species and thus are inferior for recreation, and inappropriate for the maintenance of native biodiversity (Koch and Skovsgaard, 1999; Carnus et al., 2006; Gentry et al., 2006). For example, in Hungary $20 \%$ of forest cover is the black locust (Robinia pseudoacacia), originally from North America, and $15 \%$ is black pine (Pinus nigra), a European species, but not native to

[^0]Hungary (Mátyás, 1997). There are similar patterns for other countries (e.g., EEA, 2006).

What are the effects of exotic (i.e. introduced or nonnative) trees and forests on native wildlife? Bird species usually prefer native habitats over exotic patches (Ramos, 1996; Ortega et al., 2006). Native bushes are superior foraging sites for birds (French et al., 2005). Nesting on exotic bushes in urban parks resulted in higher rate of nest failure than nesting on native bushes (Schmidt and Whelan, 1999; Borgmann and Rodewald, 2004). Exotic forest plantations can lure settling birds into such suboptimal habitat, where nest failure is higher than in native forests (Remes, 2003).

Knowledge of the effects of exotic tree plantations on birds at the community level is limited. Several studies aimed to compare communities in native and exotic forests, using simple parameters. Species richness and abundance is less variable than species composition in native versus exotic forests (Hausner et al., 2002; Johnson and Freedman, 2002; Steverding and Leuschner, 2002; Bakker and Higgins, 2003), although some studies have failed to detect any differences (Donald et al., 1998; Fleishman et al., 2003; Wilson et al., 2006). The underlying mechanism of community differences between native and exotic forests may be the predation pressure on avian broods (Barber et al., 2001; Carignan and Villard, 2002), or the selective habitat preferences of species (Lerner and Stauffer, 1998). However, we have been unable to find any comparison of the species-area relationship (SAR) in exotic versus native forests. This is surprising, because SAR is a basic rule of ecology, stating that species richness increases with increasing sample area (Rosenzweig, 1995; Báldi and McCollin, 2003; Drakare et al., 2006). Thus, the important question is whether a fundamental ecological relationship is altered by one of the most peculiar and pervasive human activities of modern times - to introduce plant species to areas outside their native ranges. In this study we investigated if exotic forests harbour fewer species than native patches. We compared the species-area relationship of bird assemblages in native versus exotic forest patches in Eastern Hungary. We hypothetized that SAR will depend on forest naturalness, and on the habitat specialization of species. Naturalness probably acts via heterogeneity, which is higher in native than in exotic forests (Thompson et al., 2003; Bartha et al., 2006). Forest patches are islands for forest specialist species, but less so for generalist species, which may occur in the surrounding landscape. This may mask the general species-area relationship (e.g. Magura et al., 2001). Therefore, we expect significant SAR in native forest patches and forest specialist bird species, and weak, if any in exotic forest patches with generalist species. Further, we expect similar species richness values in small native and exotic forest patches, where no real interior habitat is available. However, different species richnesses are expected in large patches due to difference in the species-area relationships. Such finding may have important nature conservation consequences for the maintenance of biodiversity.

## 2. Study area and methods

The study area is located in Eastern Hungary on the SzatmárBereg plain (Fig. 1) ( $\mathrm{N} 48^{\circ} 05^{\prime} 55^{\prime \prime}$, E $22^{\circ} 30^{\prime} 22^{\prime \prime}$ ). The plain is
covered by pastures and agricultural land with scattered forest patches. The natural vegetation of Central-European plains is forest, but the millennia of human activities (e.g., cattle grazing) modified it into primarily grassland areas with small forest patches (Standovár and Primack, 2001). Roughly $80 \%$ of the region is farmland. The surroundings was similar for all studied patches, grassland and/or arable fields. Native and exotic tree species were present in the patches. We chose 13 forest patches with only deciduous native tree species (Quercus robur, Carpinus betulus, Quercus petraea, Populus canescens, Populus alba, with a few Acer campestre, Salix alba and Fraxinus angustifolia), and 14 with only deciduous exotic tree species (Robinia pseudoacacia, Populus canadensis, with few individuals of Quercus rubra, Acer negundo, Salix matsudana, Amorpha fruticosa and Fraxinus pennsylvanica). The age of studied forest patches was 35-70 years. Three area categories of forest fragments were established: small ( $<10 \mathrm{ha}$ ), medium ( $10<100 \mathrm{ha}$ ), and large (>100 ha) (Table 1). Within size class fragments with natural tree composition and those with exotic trees were distinguished. There was no significant difference between the mean area of these two groups within area category ( $t_{7}=0.289, p=0.781$ for small forests, $t_{7}=$ $-0.812, p=0.443$ for medium forests, and $t_{7}=2.853, p=$ 0.064 for large forests). Forest patch heterogeneity was estimated by eye as percent cover of shrubs and number of nest holes within the censused areas.

Breeding bird communities were censused in the forest patches in the breeding season of 2003. Three censuses were carried out during the season (April, May and early June), only under good weather conditions (no wind or rain), from sun rise to 9 a.m. (Moskát, 1987). We applied a standard point census technique ( 100 m radius, 5 min census time); all birds seen or heard were recorded. Sampling effort was standardized for all patches - since many patches were only a few hectares large, one point per patch at such patches was possible to census. We distinguished between forest specialist and habitat generalist species based on literature data (Snow and Perrins, 1998), considering local conditions.

The species-area relationship was established with the most frequently used log-log transformed model (Rosenzweig, 1995), using individual patch areas (not the categories) in the calculation. The number of forest specialist and generalist bird species between the two forest types (native and exotic) with similar size were examined by repeated measure analysis of variance (ANOVA). Forest type (native or exotic) was considered as factor and the time of the counting (April, May and June) were used as repeated measures. The data were normalized by $\log (x+1)$ transformation. When the results of the ANOVA showed that there was difference in the species richness among the forest types, this was tested by a Tukey-type multiple comparison (Sokal and Rohlf, 1981). The analyses were carried out using the SPSS-PC program (SPSS, 1999).

## 3. Results

Shrub cover did not differ significantly between native and exotic forest patches $\left(t_{7}=0.324, p=0.755\right.$ for small forests, $t_{7}=0.040, p=0.969$ for medium forests, and $t_{7}=0.828$,


Fig. 1 - Location of the study area in Eastern Hungary. Triangles represent forest patches of exotic tree species and circles native tree species. The size of the mark represents three area categories of the patches ( $<10 \mathrm{ha}, 10-100 \mathrm{ha},>100 \mathrm{ha}$ ).
$p=0.435$ for large forests). There were more nest holes in native forest patches compared to the exotic patches in small and large forests, but not in the medium size category ( $\mathrm{t}_{7}=3.653, p=0.008$ for small forests, $\mathrm{t}_{7}=1.005, p=0.348$ for medium forests, and $t_{7}=2.543, p=0.038$ for large forests).

Altogether 41 bird species were observed during the censuses. Most species ( $92 \%$ ) were observed in both the native and the exotic forest patches (Appendix). The species-area relationship was significantly positive in the native forest patches for forest specialist species, and positive but not significant for generalist species (Table 2). No species-area relationship were found for bird assemblages in exotic forest patches (Table 2).

Comparing bird species richness between native and exotic forest patches of similar size classes revealed that the small
and medium patches were not significantly different for either forest specialist or generalist species. In large forests, however, bird species richness was significantly higher in native than in exotic forests for forest specialist species, but not for generalist species (Table 3). This relationship between the native and exotic forest patches seems to be stable, at least within season, because the interaction term of time * naturalness was on no occasion significant (Table 3). The time effect alone indicated significant decline in species number from April to June in the small forest patches (for both forest and generalist species), and for forest specialist species in medium forest patches. The number of generalist species did not change within season in medium-sized forest patches, and none of the groups declined over the season in the large forest patches (Table 3).

Table 1 - Characteristics of fragments and the number of observed bird species in three size groups of native and exotic forest patches in Eastern Hungary

|  | Small |  | Medium |  | Large |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Native | Exotic | Native | Exotic | Native | Exotic |
| Number of fragments | 5 | 4 | 4 | 5 | 4 | 5 |
| Area (ha) $\pm$ s.d. | $7.1 \pm 1.86$ | $6.6 \pm 2.92$ | $41.0 \pm 28.94$ | $58.7 \pm 34.99$ | $425.0 \pm 206.16$ | $130.0 \pm 18.71$ |
| Number of nestholes $\pm$ s.d. | $697 \pm 344$ | $55 \pm 49$ | $465 \pm 447$ | $204 \pm 336$ | $725 \pm 284$ | $172 \pm 351$ |
| Cover of shrubs (\%) $\pm$ s.d. | $44 \pm 27.9$ | $38 \pm 32.3$ | $33.75 \pm 31.5$ | $33 \pm 24.4$ | $46 \pm 35.0$ | $28 \pm 31.2$ |
| Number of observed forest specialist species | $12.2 \pm 1.10$ | $10.5 \pm 1.73$ | $13.8 \pm 4.99$ | $14.0 \pm 5.61$ | $16.5 \pm 1.26$ | $12.4 \pm 2.19$ |
| Number of observed generalist species | $3.2 \pm 1.64$ | $3.3 \pm 2.06$ | $1.8 \pm 1.50$ | $3.0 \pm 1.58$ | $1.8 \pm 0.96$ | $4.0 \pm 1.58$ |

Table 2 - Equations of the species-area relationship of forest specialist birds in patches of native and exotic trees in Eastern Hungary. Equations are given for forest specialist and generalist species separately. Asterisks indicate significant relationship

Equation $\quad R \quad n \quad p$
Native forests
$\begin{array}{lllll}\text { Number of forest } & \mathrm{Y}=0.82+0.09 \mathrm{X} & 0.66 & 13 & 0.02^{*}\end{array}$
specialist species
Number of generalist species $\quad \mathrm{Y}=0.46-0.01 \mathrm{X} \quad-0.07 \quad 13 \quad 0.84$
Exotic forests
$\begin{array}{lllll}\text { Number of forest } & Y=0.76+0.09 X & 0.40 & 14 & 0.16\end{array}$
specialist species
Number of generalist species $\quad \mathrm{Y}=0.47+0.03 \mathrm{X} \quad 0.10 \quad 14 \quad 0.73$

## 4. Discussion

We studied bird communities in a landscape with patches of natural and exotic tree species. The majority of the landscape was farmland (arable fields and grasslands), with scattered forest patches, including the studied patches. Although it is known that different landscape matrix influences species diversity even in similar forest patches (Lindenmayer et al., 2002; Watson et al., 2005), here we had the same landscape type as matrix. Therefore, we supposed that the matrix effect was similar for all patches, thus, did not bias the results.

There was no difference in bird species composition of natural versus exotic forest patches. In a study of birds in tree plantations versus native shrub patches in the Negev, Israel, Shochat et al. (2001) found only a small overlap of bird species. Probably the difference between exotic and native patches in our study was not large enough to exclude bird species, only to influence the frequency of occurrence. The communities showed different responses to area; the spe-cies-area relationship explained the number of bird species (i.e. significant positive species-area relationship) in native forests, but not in exotic forest patches. A similar pattern was found by Shochat et al. (2001) for native scrub fragments versus planted forests; bird species richness depended on area Q1 in the former, but not in the latter. Santos et al. (2006) compared native oak and mature pine plantations in Spain, and they found similar bird species richness in native and plantation archipelagoes, and different species-area relationship, just in this study. However, all species-area regression models were significant in their study (Santos et al., 2006).

The difference between native and exotic forests was more pronounced, if species were classified according to their specificity to the forest habitat. Species richness of generalists was not related to forest area at all. The species - area relationship of forest specialist bird species was not significant in exotic forest patches, but was significant (positive) in native patches. The clear difference between the response of specialist and generalist (including exotic) species to fragmentation was described for several taxa, including birds (McCollin, 1993; Germaine et al., 1998), plants (Abbott, 1992; Bakker and Higgins, 2003) and invertebrates (Magura et al., 2001; Ostergard and Ehrlen, 2005; Ouin et al., 2006). More generally, species traits are known to confound the SAR (Ewers and Didham, 2006).

There may be two potential mechanisms to explain the presence of SAR in native, but not in exotic forest patches. First, there is usually a basic difference between exotic and native forests in spatial heterogeneity; native forests are more heterogeneous (Mátyás, 1997; Thompson et al., 2003). We also found some indications of such a trend (more nest holes in native forests), although others were in our case not statistically significant (shrub cover). Habitat heterogeneity and species richness have a positive correlation (Tews et al., 2004), thus, the more heterogeneous and complex habitat structure of native forests may promote forest specialist bird species rather than generalist species (MacNally et al., 2000). Second, native forests are more island-like patches, because their complex structure is more different from the surrounding landscape, than the exotic forests with a simple structure. Therefore, forest specialist species are probably restricted to the native forest "isolates", with the subsequent SAR, while exotic forests might not function as isolates. This landscape effect may also be responsible for the absence of a SAR in some cases (Estades and Temple, 1999; Wethered and Lawes, 2003; Lövei et al., 2006).

Increased habitat fragmentation results in the relative increase of edge habitats due to the increase of edge/core ratio. Edges are favored by generalist and/or early successional species (Harris and Silva-Lopez, 1992; Imbeau et al., 2003; Lövei et al., 2006), therefore we expected a reverse relationship for generalist species with forest area: as forest area increases the proportion of edges, and the generalists species, is decreasing. Although these were significant neither for native ( $R=-0.1$ ) nor for exotic $(R=0.1)$ forests, the trends of the regression coefficients (Table 2) support this hypothesis.

The SAR is a fundamental rule of ecology, but it still suffers from several biases (Báldi and McCollin, 2003). Here we demonstrated the key role of native/exotic species composition, that is the quality of habitats and the specialist/generalists character of target species on the SAR (Lövei et al., 2006). An archipelago of non-native habitats, and/or non-native species in the archipelago may lead to the breakdown of the SAR.

The age of forest patches may be relevant factor in determining species richness. In this study the age of patches were $35-70$ years, that none of them were old growth. Humphrey (2005) and Santos et al. (2006), for example, showed that 100-200 year old plantations already conferring substantial benefits to many species. It is clear that for the studied forest archipelago and landscape only large ( $>100 \mathrm{ha}$ ) forest patches containing native tree species can preserve natural patterns. Such patches support forest specialist bird species during the whole breeding season. We warn, however, against using only species presence in forest patches for patch evaluation: on our study nearly all the observed species (92\%) were present in both native and exotic forest patches. Therefore, simple presence-absence survey may be misleading, because it can not identify the superior value of native forests. Other studies also highlighted the subordinate role of exotic trees and bushes for nesting, foraging and community assemblage (Schmidt and Whelan, 1999; Remes, 2003; Borgmann and Rodewald, 2004; French et al., 2005). This evidence support the conclusion that the expansion of exotic trees and bushes via forestry practice and gardening will harm natural patterns and the underlying processes, hence accelerating the decline of birds in isolated forests.

Table 3 - Results of the repeated measures ANOVA for the bird species richness in deciduous forests with similar size. Naturalness of the forests (native, exotic) comprised the factor and the time of the counting (April, May and June in 2003) were used as repeated measures. Results of the Tukey test indicate which forest category differs significantly ( $p<0.05$ ) from the other; for example 'Native > Exotic' indicates that the species richness was significantly higher in the native forests than in the exotic patches

| Variable | Source | SS | df | MS | $F$ | $p$ | Tukey posteriori test |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |


| Small forests, Number of | Within-Subjects Effects |  |  |  |  |  |
| :---: | :--- | :--- | :--- | :--- | ---: | :--- |
| forest specialist species | Time | 0.130 | 2 | 0.065 | 18.100 | 0.000 |
|  | Time $\times$ Naturalness | 0.004 | 2 | 0.002 | 0.558 | 0.585 |

Small forests, Number of
generalist species

forest specialist species


Medium forests, Number of generalist species




Large forests, Number of generalist species

## Acknowledgement

We are indebted to Duncan McCollin, Béla Tóthmérész and two anonymous referees for comments on previous versions
of the paper. The research was supported by a grant of the Hungarian Scientific Research Fund (OTKA research grant no. F61651, to T. M.), and the Directorate of the Hortobágy National Park. T. M. and A. B. were Bolyai Research Fellows of the Hungarian Academy of Sciences.

## Appendix

List of observed bird species in taxonomic order from 27 forest patches in Eastern Hungary. Total number of observations during the three censuses in 2003 is given. Snow and Perrins (1998) was followed for nomenclature. Asterisks indicate forest specialist species

|  |  | Small |  | Medium |  | Large |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Native | Exotic | Native | Exotic | Native | Exotic |
| Columbidae |  |  |  |  |  |  |  |
| Columba palumbus | Woodpigeon | 0 | 6 | 2 | 3 | 0 | 0 |
| Streptopelia turtur | Turtle Dove | 7 | 7 | 4 | 5 | 5 | 6 |
| Upupidae |  |  |  |  |  |  |  |
| Upupa epops | Hoopoe | 0 | 0 | 0 | 0 | 0 | 2 |
| Picidae |  |  |  |  |  |  |  |
| Dendrocopos major* | Great Spotted Woodpecker | 7 | 2 | 3 | 6 | 8 | 9 |
| Dendrocopos medius* | Middle Spotted Woodpecker | 0 | 0 | 0 | 0 | 1 | 0 |
| Dendrocopos minor* | Lesser Spotted Woodpecker | 0 | 0 | 2 | 1 | 0 | 0 |
| Dryocopus martius* | Black Woodpecker | 1 | 0 | 1 | 4 | 3 | 0 |
| Jynx torquilla* | Wryneck | 6 | 4 | 1 | 3 | 3 | 9 |
| Picus viridis* | Green Woodpecker | 0 | 0 | 1 | 0 | 0 | 1 |
| Motacillidae |  |  |  |  |  |  |  |
| Anthus trivialis | Tree Pipit | 6 | 4 | 3 | 12 | 2 | 2 |
| Turdidae |  |  |  |  |  |  |  |
| Erithacus rubecula* | Robin | 12 | 4 | 8 | 12 | 8 | 5 |
| Luscinia megarhynchos* | Nightingale | 10 | 18 | 4 | 13 | 3 | 6 |
| Turdus merula* | Blackbird | 14 | 9 | 11 | 12 | 20 | 8 |
| Turdus philomelos* | Song Trush | 1 | 0 | 0 | 2 | 3 | 0 |
| Sylviidae |  |  |  |  |  |  |  |
| Locustella fluviatilis* | River Warbler | 1 | 3 | 1 | 2 | 0 | 9 |
| Phylloscopus collybita* | Chiffchaff | 9 | 3 | 8 | 9 | 12 | 8 |
| Phylloscopus sibilatrix* | Wood Warbler | 6 | 5 | 6 | 7 | 12 | 1 |
| Phylloscopus trochilus* | Willow Warbler | 0 | 1 | 1 | 0 | 2 | 0 |
| Sylvia atricapilla* | Blackcap | 22 | 13 | 18 | 19 | 18 | 13 |
| Sylvia curruca | Lesser Whitethroat | 0 | 1 | 0 | 0 | 0 | 0 |
| Muscicapidae |  |  |  |  |  |  |  |
| Muscicapa striata* | Spotted Flycatcher | 0 | 0 | 1 | 2 | 2 | 1 |
| Ficedula sp.* | Flycather sp. | 0 | 0 | 0 | 0 | 1 | 0 |
| Aegithalidae |  |  |  |  |  |  |  |
| Aegithalos caudatus* | Long-tailed Tit | 0 | 0 | 1 | 1 | 0 | 0 |
| Paridae |  |  |  |  |  |  |  |
| Parus ater* | Coal Tit | 0 | 0 | 0 | 0 | 3 | 0 |
| Parus caeruleus* | Blue Tit | 3 | 2 | 3 | 6 | 11 | 4 |
| Parus major* | Great Tit | 17 | 8 | 11 | 12 | 14 | 9 |
| Parus palustris* | Marsh Tit | 0 | 0 | 0 | 0 | 0 | 1 |
| Sittidae |  |  |  |  |  |  |  |
| Sitta europaea* | Nuthatch | 1 | 0 | 4 | 2 | 3 | 2 |
| Certhiidae |  |  |  |  |  |  |  |
| Certhia sp.* | Treecreeper | 10 | 8 | 4 | 8 | 5 | 10 |
| Oriolidae |  |  |  |  |  |  |  |
| Oriolus oriolus* | Golden Oriol | 9 | 9 | 4 | 12 | 8 | 14 |
| Laniidae |  |  |  |  |  |  |  |
| Lanius collurio | Red-backed Shrike | 1 | 2 | 0 | 3 | 0 | 4 |
| Corvidae |  |  |  |  |  |  |  |
| Garrulus glandarius* | Jay | 1 | 0 | 2 | 3 | 2 | 1 |
| Sturnidae |  |  |  |  |  |  |  |
| Sturnus vulgaris | Starling | 21 | 21 | 17 | 8 | 10 | 48 |
| Passeridae |  |  |  |  |  |  |  |
| Passer montanus | Tree Sparrow | 0 | 0 | 0 | 1 | 0 | 0 |
| Fringillidae |  |  |  |  |  |  |  |
| Carduelis carduelis | Goldfinch | 0 | 2 | 0 | 1 | 0 | 2 |
| Carduelis chloris* | Greenfinch | 3 | 2 | 0 | 2 | 2 | 7 |
| Coccothraustes coccothraustes* | Hawfinch | 6 | 13 | 6 | 7 | 11 | 4 |
| Fringilla coelebs* | Chaffinch | 29 | 23 | 28 | 24 | 33 | 37 |
| Serinus serinus | Serin | 0 | 0 | 0 | 0 | 0 | 1 |


| Appendix (continued) |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Small |  | Medium |  | Large |  |
|  |  | Native | Exotic | Native | Exotic | Native | Exotic |
| Emberizidae |  |  |  |  |  |  |  |
| Emberiza citrinella | Yellowhammer | 5 | 2 | 2 | 7 | 2 | 10 |
| Miliaria calandra | Corn Bunting | 1 | 0 | 0 | 0 | 0 | 0 |

## REFERENCES

Abbott, I., 1992. Biogeography of grasses (Poaceae) on islands of southwestern Australia. Aust. J. Ecol. 17, 289-296.
Bakker, K.K., Higgins, K.F., 2003. Avian use of natural versus planted woodlands in eastern South Dakota, USA. Nat. Areas J. 23, 121-128.
Báldi, A., McCollin, D., 2003. Island ecology and contingent theory: the role of spatial scale and taxonomic bias. Global Ecol. Biogeogr. 12, 1-3.
Barber, D.R., Martin, T.E., Melchiors, M.A., Thill, R.E., Wigley, T.B., 2001. Nesting success of birds in different silvicultural treatments in Southeastern US pine forests. Conserv. Biol. 15, 196-207.
Bartha, D., Ódor, P., Horváth, T., Tímár, G., Kenderes, K., Standovár, T., Bölöni, J., Szmorád, F., Bodonczi, L., Aszalós, R., 2006. Relationship of tree stand heterogeneity and forest naturalness. Acta Silv. Lign. Hung. 2, 7-22.
Borgmann, K.L., Rodewald, A.D., 2004. Nest predation in an urbanizing landscape: the role of exotic shrubs. Ecol. Appl. 14, 1757-1765.
Carignan, V., Villard, M.A., 2002. Effects of variations in micromammal abundance on artificial nest predation in conifer plantations and adjoining deciduous forests. Forest Ecol. Manage 157, 255-265.
Carnus, J.M., Parrotta, J., Brockerhoff, E., Arbez, M., Jactel, H., Kremer, A., Lamb, D., O’Hara, K., Walters, B., 2006. Planted forests and biodiversity. J. Forestry 104, 65-77.
Dirzo, R., Raven, P.H., 2003. Global state of biodiversity and loss. Ann. Rev. Environ. Res. 28, 137-167.
Donald, P.F., Fuller, R.J., Evans, A.D., Gough, S.J., 1998. Effects of forest management and grazing on breeding bird communities in plantations of broadleaved and coniferous trees in western. England. Biol. Conserv. 85, 183-197.
Drakare, S., Lennon, J.J., Hillebrand, H., 2006. The imprint of the geographical, evolutionary and ecological context on speciesarea relationships. Ecol. Lett. 9, 215-227.
EEA, 2006. European Forest Types Categories and Types for Sustainable Forest Management Reporting and Policy. European Environment Agency, Copenhagen, Denmark.
Estades, C.F., Temple, S.A., 1999. Deciduous-forest bird communities in a fragmented landscape dominated by exotic plantations. Ecol. Appl. 9, 573-585.
Ewers, R.M., Didham, R.K., 2006. Confounding factors in the detection of species responses to habitat fragmentation. Biol. Rev. 81, 117-142.
Fleishman, E., McDonal, N., Mac Nally, R., Murphy, D.D., Walters, J., Floyd, T., 2003. Effects of floristics, physiognomy and non-native vegetation on riparian bird communities in a Mojave Desert watershed. J. Anim. Ecol. 72, 484-490.
French, K., Major, R., Hely, K., 2005. Use of native and exotic garden plants by suburban nectarivorous birds. Biol. Conserv. 121, 545-559.
Gentry, D.J., Swanson, D.L., Carlisle, J.D., 2006. Species richness and nesting success of migrant forest birds in natural river corridors and anthropogenic woodlands in southeastern South Dakota. Condor 108, 140-153.
Germaine, S.S., Rosenstock, S.S., Schweinsburg, R.E., Richardson, W.S., 1998. Relationship among breeding birds,
habitat, and residential development in Greater Tucson, Arizona. Ecol. Appl. 8, 680-691.
Harris, L.D., Silva-Lopez, G., 1992. Forest fragmentation and the conservation of biological diversity. In: Fiedler, P.L., Jain, S.K. (Eds.), Conservation Biology. The Theory and Practice of Nature Conservation Preservation and Management. Chapman and Hall, New York, USA, pp. 197-237.
Hausner, V.H., Yoccoz, N.G., Strann, K.B., Ims, R.A., 2002. Changes in bird communities by planting non-native spruce in coastal birch forests of northern Norway. Ecoscience 9, 470-481.
Humphrey, J.W., 2005. Benefits to biodiversity from developing old-growth conditions in British upland spruce plantations: a review and recommendations. Forestry 78, 33-53.
Imbeau, L., Drapeau, P., Mönkönnen, M., 2003. Are forest birds categorised as "edge species" strictly associated with edges? Ecography 26, 514-520.
Johnson, G.A.M., Freedman, B., 2002. Breeding birds in forestry plantations and natural forest in the vicinity of Fundy National Park, New Brunswick. Can. Field-Nat 116, 475-487.
Koch, N.E., Skovsgaard, J.P., 1999. Sustainable management of planted forests: some comparisons between Central Europe and the United States. New Forests 17, 11-22.
Lacaze, J.F., 2000. Forest management for recreation and conservation: new challenges. Forestry 73, 137-141.
Lerner, S.D.Z., Stauffer, D.F., 1998. Habitat selection by Blackburnian Warblers wintering in Colombia. J. Field Ornithol. 69, 457-465.
Lewis, S.L., 2006. Tropical forests and the changing earth system. Phil. Trans. Roy. Soc. B - Biol. Sci. 361, 195-210.
Lindenmayer, D.B., Cunningham, R.B., Donnelly, C.F., Nix, H., Lindenmayer, B.D., 2002. Effects of forest fragmentation on bird assemblages in a novel landscape context. Ecol. Mon. 72 (1), 1-18.
Lövei, G.L., Magura, T., Tóthmérész, B., Ködöböcz, V., 2006. The influence of matrix and edges on species richness patterns of ground beetles (Coleoptera: Carabidae) in habitat islands. Global Ecol. Biogeogr. 15, 283-289.
MacNally, R., Bennett, A.F., Horrocks, G., 2000. Forecasting the impacts of habitat fragmentation. Evaluation of speciesspecific predictions of the impact of habitat fragmentation on birds in the box-ironbark forests of central Victoria, Australia. Biol. Conserv. 95, 7-29.
Magura, T., Ködöböcz, V., Tóthmérész, B., 2001. Effects of habitat fragmentation on carabids in forest patches. J. Biogeogr. 28, 129-138.
Mátyás, C., 1997. Forestry ecology. Mezőgazda, Budapest, Hungary (In Hungarian).
McCollin, D., 1993. Avian distribution patterns in a fragmented wooded landscape (North Humberside, UK): the role of between-patch and within-patch structure. Global Ecol. Biogeogr. Lett. 3, 48-62.
Moskát, C., 1987. Estimating bird densities during the breeding season in Hungarian deciduous forests. Acta Reg. Soc. Sci. Litt. Gothoburgensis Zool 14, 153-161.
Ortega, Y.K., McKelvey, K.L., Six, D.L., 2006. Invasion of an exotic forb impacts reproductive success and site fidelity of a migratory songbird. Oecologia 149, 340-351.

Ostergard, H., Ehrlen, J., 2005. Among population variation in specialist and generalist seed predation - the importance of host plant distribution, alternative hosts and environmental variation. Oikos 111, 39-46.
Ouin, A., Sarthou, J.-P., Bouyjou, B., Deconchat, M., Lacombe, J.-P., Monteil, C., 2006. The species-area relationship in the hoverfly (Diptera, Syrphidae) communities of forest fragments in southern France. Ecography 29, 183-190.
Ozanne, C.M.P., Anhuf, D., Boulter, S.L., Keller, M., Kitching, R.L., Korner, C., Meinzer, F.C., Mitchell, A.W., Nakashizuka, T., Dias, P.L.S., Stork, N.E., Wright, S.J., Yoshimura, M., 2003. Biodiversity meets the atmosphere: a global view of forest canopies. Science 301, 183-186.
Ramos, J.A., 1996. Introduction of exotic tree species as a threat to the Azores bullfinch population. J. Appl. Ecol. 33, 710-722.
Remes, V., 2003. Effects of exotic habitat on nesting success, territory density, and settlement patterns in the Blackcap (Sylvia atricapilla). Conserv. Biol. 17, 1127-1133
Rosenzweig, M., 1995. Species Diversity in Space and Time. Cambridge University Press, Cambridge, UK.
Schmidt, K.A., Whelan, C.J., 1999. Effects of exotic Lonicera and Rhamnus on songbird nest predation. Conserv. Biol. 13, 1502-1506.
Shochat, E., Abramsky, Z., Pinshow, B., 2001. Breeding bird species diversity in the Negev: effects of scrub fragmentation by planted forests. J. Appl. Ecol. 38, 1135-1147.
Snow, D., Perrins, C., 1998. The Complete Birds of the Western Palearctic on CD-Rom. Oxford University Press, Oxford, UK.

Sokal, R.R., Rohlf, F.J., 1981. Biometry. W.H. Freeman, New York USA.
SPSS, 1999. SPSS Base 10.0. SPSS Incorporation, Chicago, USA.
Standovár, T., Primack, R.B., 2001. Essentials of Conservation Biology. Nemzeti Tankönyvkiadó, Budapest (In Hungarian).
Steverding, M., Leuschner, C., 2002. Effects of Norway spruce monocultures on the structure of bird communities in a submontane-montane forested landscape of Central Germany. Forstwiss. Centralbl. 121, 83-96.
Tews, J., Brose, U., Grimm, V., Tielbörger, K., Wichmann, M.C., Schwager, M., Jeltsch, F., 2004. Animal species diversity driven by habitat heterogeneity/diversity: the importance of keystone structures. J. Biogeogr. 31, 79-92.
Thompson, I.D., Baker, J.A., Ter-Mikaelian, M., 2003. A review of the long-term effects of post-harvest silviculture on vertebrate wildlife, and predictive models, with an emphasis on boreal forests in Ontario, Canada. Forest Ecol. Manage 177, 441-469.
Watson, J.E.M., Whittaker, R.J., Freudenberger, D., 2005. Bird community responses to habitat fragmentation: how consistent are they across landscapes? J. Biogeogr. 32, 1353-1370.
Wethered, R., Lawes, M.J., 2003. Matrix effects on bird assemblages in fragmented Afromontane forests in South Africa. Biol. Conserv. 114, 327-340.
Wilson, M.W., Pithon, J., Gittings, T., Kelly, T.C., Giller, P.S., O'Halloran, J., 2006. Effects of growth stage and tree species composition on breeding bird assemblages of plantation forests. Bird Study 53, 225-236.


[^0]:    * Corresponding author at: Tel.: +36 1210 1075; fax: +36 13342785.

    E-mail addresses: baldi@nhmus.hu, andrasbaldi@hotmail.com (A. Báldi). 1146-609X/\$ - see front matter © 2008 Published by Elsevier Masson SAS. doi:10.1016/j.actao.2007.11.007

